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Nitrite tolerance of different ectomycorrhizal and wood- and litter-decomposing fungi

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The effect of nitrite (NO_2^-) on the *in vitro* radial growth of 10 ectomycorrhizal and 11 wood- and 19 litter-decomposing fungi was studied by cultivating mycelia on malt extract (ME) agar supplemented with different amounts of NaNO_2 . Nitrite tolerance differed very much between fungus species. *Piptoporus betulinus* and *Collybia butyracea* were the only fungi, which were not able to grow on NO_2^- -media, while a mycorrhizal fungus *Cenococcum geophilum* as well as a litter-decomposing fungus *Coprinus cinereus* grew even at $2500 \text{ mg l}^{-1} \text{ NaNO}_2$. Generally, the litter-decomposing fungi thriving in gardens, dung and other nitrogen (N) rich places were the most NO_2^- -tolerant species which shows that they have adapted to continuous high supply of nitrogen. The possible mechanisms involving in NO_2^- -tolerance are discussed.

Key words: Nitrite (NO_2^-), tolerance, ectomycorrhizal fungi, wood and litter decomposing fungi.

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Introduction

Nitrite (NO_2^-) is an intermediate product in the conversion of ammonium (NH_4^+) to nitrate (NO_3^-), of NO_3^- to nitrogen gas (N_2) in soils and of NO_3^- to NH_4^+ in plants and fungi. Owing to faster oxidation of NO_2^- to NO_3^- than of NH_4^+ to NO_2^- , accumulation of NO_2^- is limited in soil, and its concentration does not normally exceed 1 mg g^{-1} soil (Haynes and Sherlock 1986). Concentration of NO_2^- may, however, dramatically increase as a result of fertilization with urea (Löffler et al. 1986, Monaghan & Barrachlough 1992) or inorganic N (Jones & Schwab 1993, Burns et al. 1995) or of industrial N deposition as well. According to Bingham et al. (1954) concentration of NO_2^- above $50 \text{ } \mu\text{g g}^{-1}$ in the root zone produces toxicity symptoms. In acidic conditions, such as in Nordic forest soils, toxic level can be reached already at much lower concentration due to forma-

tion of uncharged nitrous acid (Lee 1979, Zsolbas et al. 1993).

Reduction to NH_4^+ and further incorporation to organic form is a prerequisite for utilization of NO_2^- in organisms. Reduction of NO_2^- to NH_4^+ is catalyzed by nitrite reductase enzyme (NiR) which has been found both in chlorophyllous and non-chlorophyllous parts of plants (e.g. Duncanson et al. 1992, Wray 1993, Seith et al. 1994). Reduction of NO_2^- in fungi has not been investigated as widely as in plants, and the studies have mainly concentrated on lower fungi, such as *Neurospora crassa* (Lafferty et al. 1974, Greenbaum et al. 1978, Exley et al. 1993) and *Candida nitratophila* (Al Kubisi et al. 1996). Reduction of NO_2^- has also been observed in some wood-decomposing fungi (Gundersen 1967) as well as in an ectomycorrhizal fungus *Hebeloma*

cylindrosporum (Plassard et al. 1984). As high concentrations NO_2^- is known to be toxic to fungi (Löffler et al. 1986), and therefore it has been used to inhibit infection of, for example, *Heterobasiodion annosum* in tree stumps (Gundersen 1967, Schönhar 1997).

In order to obtain more information about NO_2^- tolerance among higher fungi, we investigated the ability of both ectomycorrhizal and wood- and litter-decomposing fungi to grow *in vitro* on the media with different concentrations of NO_2^- .

Table 1. Fungal species used in the study.

Species
Mycorrhizal fungi
<i>Amanita muscaria</i> (L. : Fr.) Hook.
<i>Amanita rubescens</i> (Pers. : Fr.) S.F. Gray
<i>Cenococcum geophilum</i> (Sow.) Ferd. & Winge
<i>Hebeloma</i> sp.
<i>Leccinum scabrum</i> (Bull. : Fr.) S.F. Gray
<i>Paxillus involutus</i> (Batsch. : Fr.) Fr.
<i>Suillus granulatus</i> (L. : Fr.) Roussel
<i>Suillus luteus</i> (L. : Fr.) Roussel
<i>Suillus variegatus</i> (Sw. : Fr.) O. Kuntze
<i>Tricholoma album</i> (Fr.) Kumm.
Wood-decomposing fungi
<i>Fomitopsis pinicola</i> (Sw. : Fr.) P. Karst.
<i>Hypholoma fasciculare</i> (Huds. : Fr.) Kumm.
<i>Hypholoma lateritium</i> (Scaeff. : Fr.) Schroet.
<i>Ischnoderma benzoinum</i> (Wahlenb. : Fr.) P. Karst.
<i>Kuehneromyces mutabilis</i> (Scaeff. : Fr.) Sing. & Smith
<i>Lentinus edodes</i> (Berk.) Sing.
<i>Lentinus lepideus</i> (Fr. : Fr.) Fr.
<i>Piptoporus betulinus</i> (Bull. : Fr.) P. Karst.
<i>Pleurotus pulmonarius</i> (Fr.) Quel.
<i>Stereum sanguinolentum</i> (Alb. & Schw.) Fr.
<i>Trametes versicolor</i> (L. : Fr.) Pilát
Litter-decomposing fungi
<i>Agrocybe molesta</i> (Lasch) Sing.
<i>Coprinus bisporus</i> Lange
<i>Coprinus cinereus</i> (Scaeff. : Fr.) S.F. Gray
<i>Coprinus disseminatus</i> (Pers. : Fr.) S.F. Gray
<i>Coprinus heptemerus</i> M. Lange & A.H. Smith
<i>Coprinus micaceus</i> A (Bull. : Fr.) Fr.
<i>Coprinus micaceus</i> B (Bull. : Fr.) Fr.
<i>Coprinus tuberosus</i> Quel.
<i>Cyathus striatus</i> (Hudson : Pers.) Willd.
<i>Collybia butyracea</i> (Bull. : Fr.) Kumm.
<i>Leucoagaricus cretaceus</i> (Bull. : Fr.) Moser
<i>Marasmius androsaceus</i> (L. : Fr.) Fr.
<i>Marasmius bulliardii</i> Quel.
<i>Marasmius scorodonius</i> (Fr. : Fr.) Fr.
<i>Mycena galopus</i> (Pers. : Fr.) Kumm.
<i>Mycena vulgaris</i> (Pers. : Fr.) Kumm.
<i>Mycena vitilis</i> (Fr.) Quel.
<i>Micromphale perforans</i> (Fr.) Sing.
<i>Tephrocye tylicolor</i> (Fr.) Moser

Materials and methods

Fungal material

All the fungi of the present study were from the culture collection of Department of Plant Biology in University of Helsinki (Table 1). The fungi, except the genus *Coprinus*, have been isolated from surroundings of Helsinki by Dr. Veikko Hintikka. The *Coprinus* species have been isolated by Mr. Paavo Höijer, Porvoo. The fungi were cultivated on 1 % malt extract (ME) Difco or Melin Norkrans (MMN) (Marx 1969) agar slants at ca. +5 °C and were transferred to fresh media about twice a year.

Experimental design

The fungi were cultivated in Petri dishes containing 1 % ME (Difco), 1 % agar (Bacteriologique type A) and different amounts of NaNO_2 (Table 2).

Table 2. NaNO_2 concentrations and pH (H_2O) of the media at the beginning of the experiment.

NaNO_2 concentration (mg l^{-1})	pH
0	5.81
25	5.90
50	5.60
100	6.12
250	6.16
500	6.28
750	6.31
1000	6.32
2500	6.40

Filter-sterilized (Millipore, pore size 0.2 μm) NaNO_2 was added to ME agar after autoclaving. The pH (H_2O) of the media was determined before cultivation. It increased from pH 5.6 (50 mg l^{-1}) NaNO_2 to 6.4 (2500 mg l^{-1} NaNO_2 , Table 2).

For the experiment, a mycelial plug, about 5 mm in diameter, was transferred to fresh medium and cultivated in the dark at 21 °C. The radial growth (mm) of the fungi was measured after the growing period of 2 (litter and wood decomposing fungi), 4 (ectomycorrhizal fungi, except *Suillus* spp.) or 7 (*Suillus* species) weeks.

Results

Responses to NaNO_2 differed remarkably between fungus species (Table 3). Although all ectomycorrhizal fungi were able to grow on the NO_2^- media, as a group they were more sensitive to NO_2^- than the wood- and litter-decomposing fungi. Within mycorrhizal fungi, *Amanita* species were the most sensitive. The growth of *A. rubescens* was almost completely inhibited even at the lowest NaNO_2 concentration (25 mg l⁻¹ NaNO_2) and that of *A. muscaria* at 50 mg l⁻¹ NaNO_2 .

Cenococcum geophilum was the exception among mycorrhizal fungi by growing at the highest concentration of NaNO_2 (2500 mg l⁻¹).

Wood-decomposing fungi formed a heterogeneous group in NO_2^- tolerance. *Piptoporus betulinus* was the only fungus, which was not able to grow on NO_2^- media (Table 3). Because it grew quite well on the control medium (no added NaNO_2), mycelium was viable. *Hypholoma fasciculare* and *Trametes versicolor* were the most tolerant species by growing at 750 mg l⁻¹ NaNO_2 . Host plant of the decomposing fungus had no correlation with NO_2^- tolerance of the species.

Generally, the litter-decomposing fungi growing in gardens, roadsides, dung and other N rich places tolerated NO_2^- best. For example, all the *Coprinus* species were able to grow at 250 mg l⁻¹ or at higher concentration of NaNO_2 . *Coprinus cinereus* would have tolerated even higher concentration than 2500 mg l⁻¹, because on this medi-

Table 3. Ability of different ectomycorrhizal and wood and litter-decomposing fungi to grow on NaNO_2 agar media. The highest concentration (g l⁻¹), at which growth was still recorded.

NaNO_2 - concentration	Ectomycorrhizal fungi ¹	Wood-decomposing fungi ²	Litter-decomposing fungi ²
0		<i>Piptoporus betulinus</i>	<i>Collybia butyracea</i>
25	<i>Amanita rubescens</i>		<i>Marasmius androsaceus</i>
50	<i>Amanita muscaria</i> <i>Paxillus involutus</i> <i>Suillus variegatus</i> <i>Suillus granulatus</i>	<i>Lentinus lepideus</i>	<i>Micronphale perforans</i> <i>Mycena vitilis</i> <i>Mycena calopus</i>
100	<i>Leccinum scabrum</i> <i>Suillus luteus</i> <i>Tricholoma album</i>	<i>Ischnoderma benzoinum</i> <i>Kuehneromyces mutabilis</i> <i>Lentinus edodes</i>	<i>Mycena vulgaris</i>
250	<i>Hebeloma</i> sp.	<i>Hypholoma lateritium</i> <i>Stereum sanguinolentum</i>	<i>Coprinus heptemerus</i> <i>Coprinus micaceus</i> A <i>Coprinus tuberosus</i> <i>Cyanthus striatus</i> <i>Marasmius bulliardii</i> <i>Marasmius scorodoni</i>
500		<i>Pleurotus pulmonarius</i> <i>Fomitopsis pinicola</i>	<i>Coprinus micaceus</i> B <i>Tephrocye tylicolor</i>
750		<i>Hypholoma fasciculare</i> <i>Trametes versicolor</i>	<i>Agrocybe molesta</i> <i>Coprinus bisporus</i> <i>Coprinus disseminatus</i> <i>Leucoagaricus cretaceus</i> <i>Coprinus cinereus</i>
1000			
2500	<i>Cenococcum geophilum</i>		

¹ 4 weeks, except *Suillus* species 7 weeks, after inoculation.

² 2 weeks after inoculation.

um mycelium covered half of the Petri dish. Fungi decomposing needle litter were sensitive to NO_2^- ; *Collybia butyracea* was not able to grow on NO_2^- media at all.

Discussion

Although NO_2^- plays an important role in N metabolism in both plants and fungi, as high concentrations it is a ion of possible toxicity. It can inhibit the uptake of some important cations (Zolbos et al. 1993) and destroy both DNA and RNA by deaminating purine rings. Inhibitory effects of NO_2^- are also related to its similarity to a sulphite ion (Pateman & Kinghorn 1976).

Ability to tolerate inorganic N, NO_2^- included, varies greatly between fungus species. This has been observed as changes in fungal population in areas polluted by N originating from industry and agriculture (e.g. Ohenoja 1978, Brandrud 1995, Holopainen et al. 1996). In the study of Holopainen et al. (1996), ectomycorrhizal *Suillus*- and *Piloderma*-species sensitive to NO_3^- and NH_4^+ were replaced by more tolerant species *Cenococcum geophilum* and *Paxillus involutus* in the soil near a pulp mill in eastern Finland. In the case of *Suillus* species and *Cenococcum geophilum*, our results *in vitro* are in agreement with those obtained by Holopainen et al. (1996). In the present study, the *Paxillus involutus* strain was, however, as sensitive to NO_2^- as *Suillus* species. Arnebrant (1994) observed great difference in N tolerance between two *Paxillus involutus* strains which shows that in addition to variation between different species, ability to use and tolerate N may vary remarkably within species. Therefore, it is difficult to make generalizations for species.

Because concentration of N in wood is low, wood-decomposing fungi would be expected to be sensitive to NO_2^- . However, Gundersen (1967) found NO_2^- tolerant species and strains by cultivating mycelia on sterilized pine stem disk. In his study, all the fungi tested were able to grow even at NaNO_2 concentration of 1000 mg l^{-1} , although the growth of, for example, *Heterobasidion annosum*, *Fomitopsis pinicola* and three *Polyporus* species was reduced significantly. The growth of *Stereum sanguinolentum* was still the same as on the control medium. Wood-decomposing fungi were, however, 10–20 times more

sensitive to NO_2^- than microfungi, such as some *Penicillium* and *Trichoderma* species. In the present study, *Fomitopsis pinicola* tolerated NO_2^- better than *Stereum sanguinolentum*, but it was not able to grow at as high NO_2^- concentration as reported by Gundersen (1967).

The concentration of N in deciduous leaves is usually higher than that in needles of conifers (Berg & Cortina 1995). Therefore, the decomposing fungi of deciduous leaves might tolerate higher N concentration than fungi decaying needles. In our *in vitro* study of NO_2^- tolerance, this tendency was observed. The most tolerant fungi were, however, the species decomposing litter in gardens, dumps and roadsides rich in N. For example, the growth of most *Coprinus* species occurring in NH_4^+ - and urea-rich conditions in nature, where accumulation of NO_2^- is obvious, was inhibited only at very high concentrations of NaNO_2 .

Although synthesis and regulation of NiR enzyme is important in adaptation to high concentrations of NO_2^- , tolerance to NO_2^- in fungi depends on several mechanisms. Some plants have been observed to oxidize NO_2^- to NO_3^- and in this way avoid accumulation of NO_2^- (Funkhouser et al. 1981, Aslam et al. 1989). In addition, NO_3^- is the ion inducing NiR enzyme (Aslam et al. 1989; Wray 1993). Fungi might have the same kind of system first to induce oxidation of NO_2^- and then to increase activity of NiR with NO_3^- formed. Nitrogen may be also stored and released gradually. Deposits rich in N have been found in vacuoles of ectomycorrhizal fungus *Cenococcum geophilum* after N-treatment (Kottke et al. 1995). Amino acids and proteins have been suggested to be predominant N forms in these deposits (Walenda & Kottke 1998), but they might also contain tiny amounts of inorganic N. On the other hand, Choudary (1993) has suggested that the yeasts tolerating high concentrations of inorganic N keep the ion in a transient state until they are able to change it to a less toxic form.

Plassard et al. (1984) showed that nitrite in medium induced NiR activity in the mycelium of *Hebeloma cylindrosporum*. Induction depended, however, on pH of the media. Acidity has an important effect on both growth and N metabolism in fungi: acidic conditions increase toxicity of NO_2^- (Gundersen 1967, Plassard et al. 1984, Chang & Chung 1988). The decrease of pH from 6 to 4 seems to be most critical. In our study,

NaNO₂ increased pH of the agar media which might compensate some toxic effects of NO₂⁻. On the other hand, before toxic concentrations of NO₂⁻, too high pH (over 6) might inhibit the growth of some ectomycorrhizal fungi originally isolated from acid soil.

Nitrous acid, formed from NO₂⁻ in acid conditions, is known to react with phenolic compounds. This reaction has been suggested as a possible mechanism for gaseous N loss from acid soil (Christianson et al. 1979, Smith & Chalk 1980). The amount of phenolic compounds, especially tannins, have been shown to increase in plants after different stress conditions (Briggs 1991, Holopainen et al. 1996), but the importance of them during N stress is still unclear. The formation of phenols could be some kind of safety measure to prevent toxic effects of nitrous acid or NO₂⁻. The possibility that phenolic compounds could prevent N accumulation in cells should be investigated.

Our study shows that although *in vitro* conditions simplify responses, relationship between N concentration in the growth site of the fungus and NO₂⁻ tolerance *in vitro* can be found. Because only one strain of a species was tested and because variation within species is known to be great, generalization of tolerance is, however, difficult. Therefore, it would be important to test more strains originating from different areas.

References

- Al Kubisi, A., Ali, A.H. & Hipkin, C.R. 1996: Nitrite assimilation by the yeast *Candida nitratophila*. – *New Phytol.* 132: 313–316.
- Arnebrant, K. 1994: Nitrogen amendments reduce the growth of extramatrical ectomycorrhizal mycelium. – *Mycorrhiza* 5: 7–15.
- Aslam, M. & Huffaker, R.C. 1989: Role of nitrate and nitrite in the induction of nitrite reductase in leaves of barley seedlings. – *Plant Physiol.* 91: 1152–1156.
- Berg, B. & Cortina, J. 1995: Nutrient dynamics in some decomposing leaf and needle litter types in *Pinus sylvestris* forest. – *Scand. J. For. Res.* 10: 1–11.
- Bingham, F.T., Chapman, H.D. & Pugh, A.L. 1954: Solution–culture studies of nitrite toxicity to plants. – *Soil Sci. Soc. Am. Proc.* 18: 305–308.
- Bowsher, C.G., Hucklesby, D.P. & Emes, M.J. 1989: Nitrite reduction and carbohydrate metabolism in plastids purified from roots of *Pisum sativum* L. – *Planta* 177: 359–366.
- Brandrud, T.E. 1995: The effects of experimental nitrogen addition on the ectomycorrhizal fungus flora in an oligotrophic spruce forest at Gårdsjön, Sweden. – *For. Ecol. Manag.* 71: 111–122.
- Briggs, M.A. 1991: Influence of herbivory and nutrient availability on biomass, reproduction and chemical defences of *Lotus corniculatus* L. – *Func. Ecol.* 5: 780–786.
- Burns, L.C., Stevens, R.J., Smith, R.V. & Cooper, J.E. 1995: The occurrence and possible sources of nitrite in a grazed, fertilized, grassland soils. – *Soil Biol. Biochem.* 27: 47–59.
- Chang, T.H. & Chung, B.K. 1988: Effects of ammonium, nitrite and calcium compounds on the germination and mycelial growth of *Phytophthora capsici* causing red pepper fruit rot. – *Kor. J. Mycol.* 16: 1–8.
- Choudary, P.V. 1993: Ability to assimilate nitrogenous oxides is limited to a few species of yeasts. – *Microbios* 74: 241–248.
- Christianson, C.B., Hedlin, R.A. & Cho, C.M. 1979: Lost of nitrogen from soil during nitrification of urea. – *Can. J. Soil Sci.* 59: 147–154.
- Duncanson, E., Ip, S.-M., Sherman, A., Dennis, W.K. & Wray, J. 1992: Synthesis of nitrite reductase in regulated differently in leaf and root of barley (*Hordeum vulgare* L.). – *Plant Science* 87: 151–160.
- Exley, G.E., Colandene, J.D. & Garret, R.H. 1993: Molecular cloning, characterization and nucleotide sequence of nit-6, the structural gene for nitrite reductase in *Neurospora crassa*. – *J. Bacteriol.* 175: 2379–2393.
- Friemann, A., Lange, M., Hachtel, W. & Brinkmann, K. 1992: Induction of nitrate assimilatory enzymes in the tree *Betula pendula*. – *Plant Physiol.* 99: 837–842.
- Greenbaum, P., Prodouze, K.N. & Garret, R.H. 1978: Preparation and some properties of homogenous *Neurospora crassa* assimilatory NADPH–nitrite reductase. – *Biochemica et Biophysica Acta* 526: 52–64.
- Gundersen, K. 1967: Nitrite as a nutrient for microfungi of the outer stem cortex of pine and spruce and its toxicity to *Fomes annosus*. – *Stud. For. Suec.* 43: 1–19.
- Haynes, R.J. & Sherlock, R.R. 1986: Gaseous losses of nitrogen. – In: Haynes, R.J. (ed.), *Mineral nitrogen in the plant soil system*: 242–302. Academic Press, Orlando.
- Holopainen, T., Heinonen-Tanski, H. & Halonen, A.: 1996: Injuries to Scots pine mycorrhizas and chemical gradients in forest soil in the environment of a pulp mill in central Finland. – *Water Air Soil Pollut.* 87: 111–130.
- Jones, R.D. & Schwab, A.P. 1993: Nitrate leaching and nitrite occurrence in a fine-textured soil. – *Soil Science* 155: 272–282.
- Kottke, I., Holopainen, T., Alanen, E. & Turnau, K. 1995: Deposition of nitrogen in vacuolar bodies of *Cenococcum geophilum* Fr. mycorrhizas as detected by electron loss spectroscopy. – *New Phytol.* 129: 411–416.
- Lafferty, M.A. & Garret, R.H. 1974: Purification and properties of the *Neurospora crassa* assimilatory nitrite reductase. – *J. Biol. Chemistry* 249: 7555–7567.
- Lee, R.B. 1979: The effect of nitrite on root growth of barley and maize. – *New Phytol.* 83: 615–622.

- Löffler, H.J.M., Cohen, E.B., Oolbekkink, G.T. & Schippers, B. 1986: Nitrite as a factor in the decline of *Fusarium oxysporum* f.sp. *dianthi* in soil supplemented with urea or ammonium chloride. – *Neth. J. Pl. Path.* 92: 153–162.
- Marx, D.H. 1969: The influence of ectotrophic mycorrhizal fungi on the resistance of pine roots to pathogenic infection I. Antagonism of mycorrhizal fungi to root pathogenic fungi and soil bacteria. – *Phytopathol.* 56: 153–163.
- Monaghan, R.M. & Barraclough, D. 1992: Some chemical and physical factors affecting the rate and dynamics of nitrification in urine-affected soil. – *Plant & Soil* 143: 11–18.
- Ohenoja, E. 1978. Mushrooms and mushroom yields in fertilized forests. – *Ann. Bot. Fenn.* 15: 38–46.
- Pateman, J.A. & Kinghorn, J.R. 1976: Nitrogen metabolism. – In: Smith, J.E. and Berry, D.R. (eds.), *The filamentous fungi* 2: 159–237. New York. Wiley.
- Plassard, C., Mousain, D. & Salsac, L. 1984. Mesure in vivo et in vitro de l'activité nitrite reductase dans les thalles de *Hebeloma cylindrosporum*, champignon basidiomycete. – *Physiol. Veg.* 22: 147–154.
- Schönhar, S. 1997: *Heterobasidion annosum* in Norway spruce stands on base-rich soils in SW-Germany – results of 30 years research. – *Allgem. Forst- und Jagdzeitung* 168: 26–30.
- Seith, B., Setzer, B., Flaig, H. & Mohr, H. 1994: Appearance of nitrate reductase, nitrite reductase and glutamine synthetase in different organs of the Scots pine (*Pinus sylvestris* L.) seedling as affected by light, nitrate and ammonium. – *Physiol. Plant.* 91: 419–426.
- Smith, C.J. & Chalk, P.M. 1980: Fixation and loss of nitrogen during transformations of nitrite in soils. – *Soil. Sci. Soc. Am. J.* 44: 288–291.
- Wallenda, T. & Kottke, I. 1998. Nitrogen deposition and ectomycorrhizas. – *New Phytol.* 139: 169–187.
- Wray, J.L. 1993: Molecular biology, genetics and regulation of nitrite reduction in higher plants. – *Physiol. Plant.* 89: 607–612.
- Zolbos, F., Haunold, E. Agnes, V. & Herger, P. 1993: Nitrite in the root zone and its effects on ion uptake and growth of wheat seedlings. – *Physiol. Plant.* 89: 626–631.

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